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Nitarsons, Inorganic Arsenic, and Other Arsenic Species in Turkey Meat: Exposure and Risk Assessment Based on a 2014 U.S. Market Basket Sample

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ABSTRACT

Background: Use of nitarsone, an arsenic-based poultry drug, may result in dietary exposures to inorganic arsenic (iAs) and other arsenic species. Nitarsone was withdrawn from the US market in 2015, but use in other countries may continue.

Objectives: Characterize the impact of nitarsone use on arsenic species in turkey meat and arsenic exposures among turkey consumers, and estimate cancer risk increases from consuming turkey treated with nitarsone prior to its 2015 US withdrawal.

Methods: Turkey from three cities was analyzed for total arsenic, iAs, methylarsonate (MA), dimethylarsinate, and nitarsone, which were compared across label type and month of purchase. Turkey consumption was estimated from NHANES to estimate daily arsenic exposures for adults and children 4–30 months of age, and cancer risks among adult consumers.

Results: Turkey meat from conventional producers not prohibiting nitarsone use showed increased mean levels (in $\mu\text{g kg}^{-1}$) of iAs (0.64) and MA (5.27) compared to antibiotic-free and organic meat (0.39 and 1.54, respectively) and meat from conventional producers prohibiting nitarsone use (0.33 and 0.28, respectively). Samples with measurable nitarsone had the highest mean iAs and MA (0.92 and 10.96, respectively). Nitarsone was higher in October samples as compared to March, possibly due to increased summer use. Based on mean iAs concentrations in samples from conventional producers with no known policy versus policies prohibiting nitarsone, estimated lifetime daily consumption by an 80kg adult, and a recently-proposed cancer slope factor, we estimated that use of nitarsone by all turkey producers would result in 3.1 additional cases of bladder or lung cancer per 1,000,000 consumers.

Conclusions: Nitarsone use can expose turkey consumers to iAs and MA. Our study supports the FDA's removal of nitarsone from the US market, and further supports its removal from the global marketplace.

INTRODUCTION

Arsenic-based drugs have been used in the production of chickens, turkeys and swine in the US since the 1940s (Silbergeld and Nachman 2008). The approvals for three of these drugs, roxarsone, arsanilic acid, and carbarsone, were withdrawn by the US Food and Drug Administration (FDA) on September 30, 2013, rendering their domestic sale illegal (United States Food and Drug Administration 2013). The approval of a fourth drug, nitarsonsone ((4-nitrophenyl)arsonic acid, $C_6H_6AsNO_5$), used in chickens and turkeys, was withdrawn by FDA in December of 2015, terminating the domestic sale of the drug.

Evidence shows that the use of arsenic-based drugs in food animal production results in human dietary exposures to arsenic, including inorganic arsenic (iAs) (Lasky et al. 2004; Wallinga 2006; Nachman et al. 2013; Liu et al. 2015a; Liu et al. 2016), and in environmental distribution of arsenic in manure (Jackson and Bertsch 2001; Bednar et al. 2003; Garbarino et al. 2003; Jackson et al. 2003; Rutherford et al. 2003; Nachman et al. 2005; Nachman et al. 2008). Previous research on arsenic-based drugs has primarily considered roxarsone, a drug used in up to 90% of domestic chicken production prior to its removal from the US market (Nachman et al. 2012). In a US-based market-basket study, we found significant increases in iAs concentrations in chicken meat from animals likely raised with roxarsone compared to meat from organic and antibiotic free chickens not fed roxarsone (Nachman et al. 2013).

Little is known, however, about the potential arsenic exposure due to nitarsonsone use in turkey production. Its clinical indication is for the prevention of blackhead disease in poultry, caused by the protozoan species *Histomonas meleagridis* (McDougald 2005). To our knowledge, no study has evaluated the distribution of arsenic species in meat from nitarsonsone-treated turkeys.

Given the similarity of roxarsone and nitarsone (Figure 1), nitarsone use could result in similar dietary exposures to arsenic for turkey consumers.

The level of nitarsone dosage during its preventive use, and the lifespan of turkeys raised for meat consumption suggest that arsenic species could accumulate in commonly-consumed tissues such as muscle, fat and skin. Although information regarding pharmaceutical use in animal production is not made public, statements from industry trade groups suggest that nitarsone was widely used in US turkey production prior to its withdrawal from the market in 2015, mostly during the first few weeks of the birds' lives and more heavily during summer months, in turkeys that will be consumed during the fall and winter (Aubrey 2013; Strom 2013).

The purpose of this study was to examine the potential impact of nitarsone use on arsenic species exposure among turkey consumers. Specifically, we studied retail turkey products to characterize the occurrence of arsenic species in meat. Given industry statements about the seasonality of nitarsone usage in turkey production, we examined differences in occurrence of arsenic species in meat from turkey products purchased in two different seasons. Using turkey consumption data from the National Health and Nutrition Examination Survey (NHANES), we estimated the lifetime average daily exposure to these arsenic species among turkey consumers, and estimated cancer risks associated with ingestion of iAs attributable to arsenical drug use in animal production.

METHODS

Sample collection and preparation

A total of 184 turkey samples were included in our study, including products from 14 producers representing 64% of the 2014 US turkey market (National Turkey Federation 2016).

Turkeys reach market weight and are processed at 4 – 5 months of age. Given this, we examined 128 samples of raw turkey meat purchased in March (to represent turkeys whose lifespan primarily did not overlap with summer months when nitarsone use may have been less likely or reduced), and 56 raw turkey samples purchased in October (to represent turkeys whose lifespan largely overlapped with the summer months when nitarsone use was more likely).

Samples purchased in March 2014 were purchased from retail grocers in three geographically diverse US metropolitan areas (Baltimore, Maryland; Denver, Colorado; and Los Angeles, California), selected to cover the East Coast, West Coast, and Midwest. We visited 12 unique stores (9 supermarket chains) in Los Angeles, 10 unique stores (6 supermarket chains) in Denver, and 12 unique stores (10 supermarket chains) in Baltimore, for a total of 36 unique stores representing 23 supermarket chains. All samples collected in October 2014 (n=56) were purchased in Baltimore, Maryland; 10 additional samples were purchased in October (as compared to Baltimore samples from March, n=46) to increase the study sample size for that month. There was slight variation in grocery chains and stores visited in March and October, but overall, there was considerable overlap in purchase sites.

All samples were taken from packages of raw turkey meat (ground or whole cuts). Within each store, we purchased two packages of a single type (ground or whole cuts) of turkey for each brand available (multiple brands were purchased when available within a store), including store-branded products. We also purchased conventionally-produced products and USDA Organic and antibiotic-free-labeled turkey products of each type and brand, when possible. If only one package was available for a type and brand of turkey, then one single package was purchased. For example, from a given store in one of the study cities, based on available products, we purchased two packages of conventional ground turkey and two packages of conventional whole

cut turkey from Producer B, and two packages of conventional ground turkey from Producer A. In this particular store, neither antibiotic-free nor USDA Organic turkey products were available.

The mean (SD) meat weight in the purchased package was 0.52 (0.15) kg for ground meat and 0.69 (0.28) kg for whole meat. From each package, 60-70 g aliquots of meat were collected and processed as described below for arsenic analyses. All whole pieces of meat examined were exclusively breast tissue. Ground meat, however, was often a mixture of breast and other unspecified muscle tissue. We did not purchase cooked products, deli meats, or products such as hot dogs, sausages, or frozen meals containing turkey because mixed ingredients in the packages could interfere with sample analysis. For example, some packages of turkey deli meat contain rice starch or carrageenan (e.g., an edible algae) used for gelling and thickening, which could have been a source of added arsenic.

Samples were prepared for freeze-drying as previously described (Nachman et al. 2013). Briefly, a roughly 60 - 70 g sub-sample of raw meat was removed from each sample package. The sub-sample was > 10% of the package weight (mean package weight was 560 g), which we assumed to be representative of the entire package. Sub-samples were individually homogenized in a food processor with the addition of 75 mL MilliQ water (Millipore Corporation) to aid blending. Blended samples were weighed and stored in sealable bags at 20 °C. Between samples, the food processor and all laboratory equipment were cleaned with hot water, soaked for 30 min in a 10% nitric acid bath, and rinsed with MilliQ water. Frozen sample homogenates were shipped on dry ice to Oregon State University for freeze-drying (SP Scientific). Sample-specific water loss factors were recorded. Freeze-dried samples were stored as a crumbled powder in 50 mL polypropylene tubes at 25°C and shipped from Johns Hopkins University to the Institute of Chemistry-Analytical Chemistry, University of Graz, Austria for arsenic analyses.

Determination of total arsenic and arsenic species

Detailed descriptions of laboratory reagents, standards, and reference materials; instrumentation; extraction of arsenic species; and HPLC-ICPMS analyses are provided in Supplemental Material. In brief, the total arsenic contents of the freeze-dried turkey samples were determined by using inductively coupled plasma mass spectrometry (ICPMS – Agilent 7900 ICPMS from Agilent Technologies) following microwave-assisted acid mineralization in an UltraCLAVE III microwave system (MLS GmbH). Arsenic standards in the range $0.01 - 20 \mu\text{g As L}^{-1}$ were used for external calibration; germanium ($20 \mu\text{g L}^{-1}$ in final solution) was added to standards and samples to normalize matrix effects. As there is currently no turkey or chicken meat reference material certified for total arsenic content, for quality control, we used the standard reference material ERM-BC211 rice (European Commission 2012), with a certified arsenic content of $260 \pm 13 \mu\text{g As kg}^{-1}$; we obtained $265 \pm 18 \mu\text{g As kg}^{-1}$ ($n=39$) over the course of the study. The turkey samples were analyzed in duplicate for their total arsenic content; when duplicate values differed by $>10\%$, which occurred with just 4 of the 184 samples, the sample was re-analyzed. Taken together, 432 total arsenic measurements (including duplicate measurements for the 184 samples, four additional measurements for outliers, and duplicate measurements for 30 blinded duplicate quality control samples) were made.

Arsenic species were determined in duplicate analysis in alkaline aqueous extracts of the freeze-dried turkey samples by using anion-exchange HPLC (high performance liquid chromatography) coupled to an ICPMS which served as arsenic-selective detector. When the values for arsenic species in the duplicates differed by more than 20% , the sample was re-analyzed. Separation of dimethylarsinate (DMA), methylarsonate (MA), iAs, and nitarsonate was achieved by gradient elution using ammonium carbonate buffer with a Dionex AS15A anion

exchange column. For quality control, we used the standard reference material ERM-BC211 (rice, certified contents of DMA $119 \pm 13 \mu\text{g As kg}^{-1}$ and iAs $124 \pm 11 \mu\text{g As kg}^{-1}$); we obtained $116 \pm 4 \mu\text{g As kg}^{-1}$ for DMA and $99 \pm 4 \mu\text{g As kg}^{-1}$ for iAs (n=9). Detection limits, on a dry weight basis, were 1, 1-2, 1, 1, and $1 \mu\text{g kg}^{-1}$ for total arsenic, iAs, MA, DMA and nitarsone, respectively.

Sample-specific concentration estimates were derived by taking the average of the replicate measurements (2-3 replicates for total arsenic and each arsenic species). For total arsenic, iAs, MA and DMA measurements below the detection limit, we imputed the value of the detection limit divided by the square root of two. For nitarsone, samples below the detection limit were given the value of 0. For each arsenic species, wet weight sample concentrations were calculated for each sample by multiplying its dry weight concentration by the sample-specific water loss dilution factor.

Samples were analyzed in a random order, and the lab was blinded to sample identity. Thirty blinded duplicate samples were analyzed separately to evaluate the performance of the method, but were not included as additional samples in the main analysis.

Other variables

We categorized samples into two groups based on their package labels to compare USDA Organic or antibiotic-free to conventionally produced samples. It is critical to note that although the USDA Organic certification program explicitly prohibits the use of arsenic-based pharmaceuticals in animal production (United States Department of Agriculture 2016b), the label “no antibiotics added” only restricts the use of antibiotics, and does not specifically prohibit arsenical antimicrobials (United States Department of Agriculture 2015). In addition, we

classified each of the 117 conventional samples (not labeled as USDA Organic or as antibiotic free) as being a conventional producer with a stated policy against nitarsone use (1 producer, 30 samples) or from a conventional producer with no known policy against nitarsone use (8 producers, 87 samples) based on information from company websites or responses to emails or phone calls requesting this information. . After laboratory analyses, all samples were further categorized based on the presence of measurable nitarsone above the detection limit.

Statistical analyses

Statistical analyses were performed with Stata 14. Mean arsenic species concentrations and 95% CIs were calculated to evaluate differences among categories of turkey samples (by package label, by nitarsone policy, by positive nitarsone detection, and by season of purchase). The seasonal analysis was also restricted to Baltimore-only in a sensitivity analysis. Pearson's coefficients were used to assess correlations between concentrations of total arsenic and arsenic species including nitarsone. Statistical significance was two-tailed and set at $\alpha=0.05$.

Estimation of population turkey consumption rates and body weights

Intake rate calculation

Population intake rates for turkey meat were derived from the dietary recall component of the 2003-2010 cycles of the National Health and Nutritional Examination Survey (NHANES) using the “survey” package in R to account for NHANES' complex survey design and sampling weights (Lumley 2014); detailed survey and dietary recall methods are available through the National Center for Health Statistics (National Center for Health Statistics 2014). Each reported food item is linked to a systematic, 8-digit U.S. Department of Agriculture (USDA) food code

and recorded in grams. Because USDA food code items often contain multiple food components (e.g. turkey sandwich), we linked participant information with the Environmental Protection Agency's (EPA) Food Commodity Index Database (FCID), developed by Office of Pesticide Programs (OPP) for use in pesticide risk analysis (Joint Institute for Food Safety and Applied Nutrition 2010). The FCID converts the total weight of each USDA food code item into the weight attributable to each commodity (e.g., "Turkey, meat"). Each commodity weight is summed across all USDA food items and expressed in $\text{g kg bodyweight (BW)}^{-1} \text{ day}^{-1}$. Because the FCID database was updated in 2005, we merged the original database for 2003-2004 participants with the updated database for 2005-2010 participants. We derived the consumer-only intake rate from any participant reporting consumption of a USDA food code item containing the FCID commodity code 5000382000, "Turkey, meat." The commodities "Turkey, skin," "Turkey, fat," "Turkey, liver," and "Turkey, meat byproducts," do not contribute to our intake rate calculations, as "Turkey, meat" best corresponds to the turkey breast meat analyzed in our study. This intake rate represents the average consumption of turkey meat, in $\text{g kgBW}^{-1} \text{ day}^{-1}$, of all individuals who reported any turkey consumption on either day of dietary recall (24.1%).

The consumer-only intake rate for turkey in baby food was derived similarly among participants age 4-30 months using the FCID commodity code 5000382001, "Turkey, meat-babyfood." Because the small number of participants consuming the commodity resulted in stratum with one primary sampling unit, we conservatively centered single-primary sampling unit stratum at the grand mean (Lumley 2014).

Among the approximately 24% of the US population that consumes "Turkey, meat," intake rates were estimated at 0.49 (95% CI: 0.47, 0.51) $\text{g kgBW}^{-1} \text{ day}^{-1}$; intake rates of "Turkey,

meat- babyfood” were estimated at 1.86 (95% CI: 1.50, 2.23) g kgBW⁻¹ day⁻¹ for children aged 4-30 months.

Body weight estimation

The adult body weight distribution for the risk assessment of “Turkey, meat” was derived from all participants 18 years and older, and the body weight distribution for “Turkey, meat- babyfood” was derived from all participants age 4-30 months.

Exposure and risk analyses

To estimate the arsenic species exposure burden attributable to turkey consumption in adults, we estimated the lifetime average daily doses (LADD) for each arsenic species for each of the categories of turkey products (by package label, by nitarsone use policy and by positive nitarsone detection) using the formula:

$$\text{LADD} = ([\text{As}] \times \text{IR}) / \text{BW}, \quad [1]$$

where LADD= lifetime average daily dose (in mg kgBW⁻¹ day⁻¹), [As] = mean arsenic species concentration (in mg kg⁻¹) in the specific category of turkey product, IR = per capita turkey intake rate and BW = body weight (80 kgBW) (United States Environmental Protection Agency 2011).

We also estimated average daily doses of arsenic species for children between the ages of 4 and 30 months that consume baby food made from turkey using the formula:

$$\text{ADD} = ([\text{As}] \times \text{IR}) / \text{BW}, \quad [3]$$

Where ADD = average daily dose (in mg kgBW⁻¹ day⁻¹), [As] = mean arsenic species concentration (in mg kg⁻¹), IR = turkey intake rate (0.019 kg day⁻¹) and BW = body weight (11.01 kgBW).

For iAs, we also estimated the difference in iAs intake for consumers of turkey produced with nitarsone as compared to consumers of turkey produced without nitarsone by subtracting the mean iAs concentration for the antibiotic-free or USDA-certified Organic group from the mean iAs concentration for the conventional with no known arsenical policy group. The resulting value is an estimate of the added iAs in the meat from the use of nitarsone. The calculation of the LADD and risk using this value reflects an estimation of the added iAs exposure and excess lifetime cancer risk attributable to the decision to use nitarsone in turkey production. We then estimated cancer risk in turkey consumers by multiplying this LADD by the Environmental Protection Agency (EPA) Integrated Risk Information System (IRIS) cancer slope factor (q^*) for iAs:

$$\text{Risk} = \text{LADD}_{\text{iAs}} \times q^* \quad [2]$$

The EPA IRIS toxicological review of inorganic arsenic is currently under reassessment and review by the National Research Council; consequently, we used a value of $25.7 \text{ [mg kgBW}^{-1} \text{ day}^{-1}]^{-1}$ proposed in a draft version of the toxicological review in 2010, reflecting the Agency's most recently published analysis of the epidemiologic literature and corresponding to lung and bladder cancers (United States Environmental Protection Agency 2010). Toxicity metrics were unavailable for other species. Population cancer burdens were calculated by multiplying the estimated cancer risk by the fraction (24%) of the 2015 US population consuming turkey (76,821,806).

We conducted sensitivity analyses to examine how the relative market share of each turkey producer would impact the estimated population cancer burden attributable to turkey consumption in the US, if nitarsone had not been withdrawn from the market. Using the most recent (2014) producer-specific data (Table S1) from the National Turkey Federation (National

Turkey Federation 2016), a trade association for the US turkey industry, we calculated the relative market share for each producer by dividing its 2014 production quantity by the production quantity summed across all producers. We then multiplied this market share by the mean iAs value (from our sampling data) for each producer (for turkey producers for whom we did not test samples, we imputed the average iAs value from the study [$0.50 \mu\text{g kg}^{-1}$]), and then summed those values to create a weighted mean iAs concentration for turkey products in the US ($0.72 \mu\text{g kg}^{-1}$). We then used this weighted mean iAs in the previously described models to estimate LADD, risk, and annual population cancer burden.

A second approach was considered to allow for comparisons between producers who likely use nitarsone and those who do not. Considering only turkey producers for whom we have samples (representing 64% of the US turkey market), we used the 2014 producer-specific data from the National Turkey Federation to calculate the relative market share for each producer by dividing its 2014 production quantity by the production quantity summed across the producers included in our study. We then categorized producers into a nitarsone positive group (including producers with at least one sample with measurable nitarsone) and a nitarsone negative group (including producers with no measurable nitarsone in any sample). Then, weighted mean iAs concentrations were calculated for the nitarsone positive ($0.74 \mu\text{g kg}^{-1}$) and negative ($0.11 \mu\text{g kg}^{-1}$) groups, separately. The difference in those weighted means ($0.63 \mu\text{g kg}^{-1}$) was considered the iAs attributable to nitarsone use. It was then used in the previously described models to estimate LADD, risk, and annual population cancer burden.

RESULTS

Arsenic species in turkey meat

The mean total arsenic concentration in turkey meat was 11.2 (95% CI: 7.2, 15.1) $\mu\text{g kg}^{-1}$ (Table 1). Mean iAs, MA, DMA and nitarsone concentrations were 0.5 (95% CI: 0.4, 0.6), 3.1 (95% CI: 2.0, 4.2), 2.4 (95% CI: 2.1, 2.6), and 0.3 (95% CI: 0.1, 0.4) $\mu\text{g kg}^{-1}$, respectively. Nitarsone was detected in 17% of samples.

By package label, conventionally-produced samples (which come from turkeys permitted to receive nitarsone) had higher levels of total arsenic ($p=0.026$), iAs ($p=0.041$), and MA ($p=0.036$) concentrations compared to the combined USDA organic and antibiotic-free labeled samples. Measurable nitarsone was observed in 21% of conventional samples, as compared to 10% in the combined antibiotic-free and USDA Organic samples. Within the conventional group, further differences were observed when samples were grouped by whether producers had a known policy or no known policy prohibiting arsenic use. Samples from conventional producers with no known policy ($n = 87$ samples) had significantly higher mean concentrations for all arsenic species than samples from conventional producers with prohibitory policies ($n = 30$ samples, $p < 0.05$ for all comparisons). Nitarsone was not found in samples from conventional producers with prohibitory policies, but was detected in 28% of samples from producers without known policies.

The greatest differences in arsenic species concentrations were observed between samples with and without measurable nitarsone ($n = 31$ and 153 samples, respectively, $p < 0.01$ for all comparisons) (Table 1). The largest difference was for MA; with mean 11.0 (95% CI: 5.0, 16.9) $\mu\text{g kg}^{-1}$ for samples with detectable nitarsone compared to 1.5 (95% CI: 1.2, 1.8) $\mu\text{g kg}^{-1}$ for samples without detectable nitarsone ($p < 0.001$).

Seasonal comparisons were examined using only Baltimore samples. A significant difference in iAs (in $\mu\text{g kg}^{-1}$) was observed ($p=0.009$) between samples purchased in March (0.40

[95% CI: 0.22, 0.60]) and October (0.79 [95% CI: 0.57, 1.02]). Nitarsons concentrations were 0.65 (95% CI: 0.12, 1.19) $\mu\text{g kg}^{-1}$ for samples purchased in October and 0.11 (95% CI: 0.04, 0.19) $\mu\text{g kg}^{-1}$ for samples purchased in March ($p=0.07$). Differences were not observed for the other species (p -values were all > 0.1), although mean total arsenic and MA concentrations were higher for samples purchased in October.

Nitarsons residues were measured in at least one sample from seven of the 19 producers included in the study (producers with three or fewer samples were grouped into an “other” category); among those, the frequency of nitarsons detection varied considerably, from 7% to 100%. The producer with a rate of 100% nitarsons measurement was part of the “other” category and accounted for two samples; none of the remaining producers ($n=8$) in the “other” category had measurable nitarsons. Among producers with measurable nitarsons, mean values ranged from 0.05 to 1.09 $\mu\text{g kg}^{-1}$.

We estimated Pearson correlation coefficients between arsenic species (Table S2). Moderate to strong correlations were observed for all measured arsenic species, with the weakest correlation being between iAs and DMA (0.54).

Exposure and risk analyses

Estimates for LADD of arsenic species by sample characteristics are presented in Table 2. Consuming turkey raised under different management regimens changed the estimated daily exposures to arsenic species and total arsenic. For example, for adults and children, individual and combined arsenic species exposures were more than 2 times higher from turkeys raised by conventional producers with no nitarsons policies compared to conventional producers with policies banning its use. For the conventional, conventional with no known arsenical policy, and

positive nitarsone detection sample categories, MA exposure accounted for the largest contribution for a single species; for the other categories, DMA contributed the most to exposure.

By group, iAs exposures were higher among samples from conventional producers without policies prohibiting nitarsone use and samples with measurable nitarsone, and lower for samples from the combined antibiotic-free and USDA Organic group and conventional producers where use was prohibited by company policy. Similar patterns were seen for average daily doses for children (Table 2). One notable difference, however, is that, considering their differing intake rates and body weights, exposures among children aged 4 – 30 months were almost 4 times higher than those for adults (e.g. when considering turkey with a positive nitarsone detection, average daily iAs doses for children aged 4 – 30 months were $1.7 * 10^{-6} \text{ mg kgBW}^{-1}\text{day}^{-1}$, as compared to $4.5 * 10^{-7} \text{ mg kgBW}^{-1}\text{day}^{-1}$ for adults).

Compared to consumers of antibiotic-free and organic turkey, we estimated that an average 80 kg person consuming $0.039 \text{ kg turkey day}^{-1}$ from conventional producers without policies prohibiting nitarsone use (mean iAs concentration of $0.00064 \text{ mg kg}^{-1}$) would ingest an additional $0.01 \text{ } \mu\text{g}$ iAs per day, resulting in a LADD of $1.22 * 10^{-7} \text{ mg kgBW}^{-1} \text{ day}^{-1}$. Based on the EPA's proposed cancer slope factor for iAs of $25.7 (\text{mg kgBW}^{-1}\text{day}^{-1})^{-1}$ (United States Environmental Protection Agency 2010), this lifetime average daily exposure would result in approximately 3.1 additional cases of bladder or lung cancer per 1,000,000 persons. Our estimates for the 2015 US population (United States Census Bureau 2014), 24% of whom are estimated to be turkey consumers, suggest that if turkey industry use of nitarsone had not been discontinued, 241 additional cases of cancer might have occurred in the US over 70 years ($3.4 \text{ cancers yr}^{-1}$ in the US population as a whole). This scenario represents the estimated increase in cancer cases if nitarsone was used in all domestically-produced turkey in the US and people were

exposed throughout their lives, compared with the numbers of cases expected in the absence of any nitarsone use by turkey producers in the US.

We conducted additional sensitivity analyses to consider the influence of the market share of turkey producers on cancer burden estimates, had nitarsone not been withdrawn. Using 2014 producer-specific data from the National Turkey Federation (National Turkey Federation 2016), we estimated the market share-adjusted population cancer burden from iAs due to consumption of turkey. Under this scenario, which employed market share-weighted iAs producer-specific concentration estimates (and the mean iAs value from our study imputed for producers we did not sample), we estimated an iAs LADD of 3.53×10^{-7} mg kgBW⁻¹ day⁻¹, a risk of 9.07×10^{-6} , and an annual population cancer burden of 10 cases yr⁻¹ attributable to turkey consumption in the United States (including contributions from both nitarsone users and non-users). To refine this estimate of burden, we estimated the market share-adjusted population cancer burden from iAs attributable to nitarsone use by separating producers into nitarsone use and non-use categories based upon our data (producers not sampled in our study were excluded from this analysis). Under this scenario, we estimated a LADD of 3.09×10^{-7} mg kgBW⁻¹ day⁻¹, a risk of 7.93×10^{-6} , and an annual population cancer burden of 9 cases yr⁻¹. As compared to the original burden estimation, which did not consider producer market share, burden estimates under these two modeled scenarios were higher. The reason for this is the uneven market share across producers; for example, the producer with the largest market share was a nitarsone user with the highest mean iAs concentration.

Since cancer slope factors are not available for other arsenic species, the contribution of MA and DMA to lung and bladder cancer risks cannot be estimated, despite evidence supporting their potential carcinogenicity (International Agency for Research on Cancer 2004).

DISCUSSION

Prior to their withdrawals in 2014 and 2015 (United States Food and Drug Administration 2015), arsenic-based drugs had been used in the US since the 1940s (Silbergeld and Nachman 2008). Our results support the hypothesis that nitarsone use increases inorganic and methylated arsenic species concentrations in turkey meat, resulting in a source of arsenic exposure for consumers of treated birds. We found that estimated arsenic exposure differed according to month of purchase, consistent with an effect of seasonal patterns of nitarsone on residues. This finding is consistent with an analysis of NHANES data collected prior to withdrawal of arsenical medications for use in poultry, which suggested that seasonal variation in turkey consumption contributed to seasonal variation in urinary arsenic concentrations in the US (Nigra et al. 2015). In addition, we found that MA concentrations were statistically elevated in samples where nitarsone was measured above the detection limit (as compared to samples where nitarsone was not found above the detection limit). Epidemiologic data describing the biological significance of MA are limited. Experimental research suggests, however, that trivalent forms of these species, which are products of methylation processes in human metabolism of arsenic (Aposhian et al. 2004), have been reported to be genotoxic (Styblo et al. 2000; Mass et al. 2001; Kligerman et al. 2003) and in mice, maternal exposures during gestation have been shown to elicit tumors at multiple sites (uterine, ovarian and adrenal tumors in females, and hepatocellular, adrenal and lung in males) in offspring (Tokar et al. 2012). Because epidemiologic investigations have focused on drinking water (where iAs dominates), little is known about the health significance of oral exposure to methylated species, though concerns have been raised about the

potential for their carcinogenicity in humans by IARC and others (International Agency for Research on Cancer 2004; deCastro et al. 2014).

Our results come soon after FDA's withdrawal of the marketing approval of nitarsonsone in the US (United States Food and Drug Administration 2015). The basis for the FDA's commitment to withdraw nitarsonsone's approval was the body of research surrounding roxarsone (United States Food and Drug Administration 2011; Conklin et al. 2012; Nachman et al. 2013; Peng et al. 2014; Liu et al. 2015b), as no data specific to nitarsonsone have been available. Our analyses provide support for FDA's precautionary action. In addition to mitigating dietary arsenic exposure, ending domestic nitarsonsone sales will eliminate the introduction of arsenic into the turkey production manure stream, removing a significant contribution of arsenic from the US food production cycle.

The FDA action for arsenical drugs in the US does not impact their use in other countries, nor does it prevent US-based pharmaceutical companies from selling these drugs in other countries. For example, arsenicals are approved for use in China, and can be found in commercial feed for poultry and swine and animal manure in China (Yao et al. 2013; Huang et al. 2014; Huang et al. 2015). The USDA is close to approving exports of cooked poultry products from China, which would be raised, slaughtered and processed in China, for the US market (United States Department of Agriculture 2016a). If these imports occur, US consumers may face arsenical exposures from poultry meat. Independent of US concerns, Chinese consumers also face health risks from consuming poultry and swine fed arsenical drugs. To avoid unnecessary exposures to arsenic species, we encourage the withdrawal of arsenical drugs globally through revision to the FAO/WHO Codex Alimentarius.

Our assessment did not specifically address non-cancer health concerns linked to chronic iAs exposure (National Research Council 2014), and little is known about hazards related to routine exposure to other arsenic species. In addition, our earlier work on roxarsone suggested that cooking changed the species profile, leaving methylated species intact but significantly decreasing the parent compound in favor of production of iAs (Nachman et al. 2013). It is unknown if cooking also affects nitarsone, but if cooking has an effect similar to its apparent effect on roxarsone, we may have underestimated potential iAs exposures to consumers of nitarsone-treated turkey meat because our estimates were based on measured concentrations in raw turkey.

We were unable to account for other potential sources of arsenic that could be related to differences in the impact of nitarsone use on measured arsenic species, including arsenic in the drinking water used by turkeys, or non-arsenical drug-related sources of arsenic in the feed used for turkeys.

Future work (much of which is dependent on the development of quantitative toxicity metrics for non-cancer health outcomes for iAs and cancer potency factors for other arsenic species) is needed to contextualize other aspects of the current and historic health burden imposed on turkey consumers by the use of nitarsone.

CONCLUSIONS

Our study provides evidence that use of nitarsone in turkey production can contribute to iAs and methylated arsenic species exposure among turkey consumers. Our findings support the US FDA's precautionary decision to withdraw nitarsone's marketing approval in 2015, and

support similar actions by other governments and international agencies to protect public health in all populations.

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Table 1. Mean (95% CI) of concentrations of total arsenic and arsenic species (in $\mu\text{g As kg}^{-1}$) in turkey meat by sample characteristics

Turkey sample classification	n	Total arsenic mean (95% CI)	iAs	Arsenic species mean (95% CI)		Nitarsonsone (% positive)	Nitarsonsone
				MA	DMA		
All	184	11.18 (7.24 - 15.11)	0.50 (0.41 - 0.59)	3.10 (1.99 - 4.21)	2.37 (2.10 - 2.64)	17%	0.27 (0.10 - 0.43)
Package label							
Conventional	117	14.53 (8.42 - 20.64)*	0.56 (0.43 - 0.70)	3.99 (2.27 - 5.72)*	2.49 (2.12 - 2.87)	21%	0.39 (0.13 - 0.65)
Antibiotic-free or organic	67	5.32 (4.33 - 6.32)	0.39 (0.35 - 0.43)	1.54 (1.21 - 1.87)	2.16 (1.80 - 2.52)	10%	0.054 (0.014 - 0.094)
Producer arsenical policy							
Conventional with no known policy	87	19.20 (11.18 - 27.21)**	0.64 (0.46 - 0.82)*	5.27 (3.01 - 7.54)*	3.09 (2.65 - 3.53)***	28%	0.53 (0.18 - 0.87)
Conventional with prohibitory policy	30	0.98 (0.82 - 1.15)	0.33 (0.29 - 0.37)	0.28 (0.24 - 0.32)	0.76 (0.65 - 0.86)	0%	NA
Nitarsonsone detection							
Negative	153	5.15 (4.40 - 5.91)***	0.42 (0.36 - 0.47)***	1.51 (1.23 - 1.78)***	1.94 (1.74 - 2.14)***	0%	NA
Positive	31	40.89 (19.97 - 61.81)	0.92 (0.51 - 1.33)	10.96 (5.01 - 16.90)	4.52 (3.51 - 5.52)	100%	1.59 (0.72 - 2.47)
Month of purchase							
March (Baltimore only)	46	7.58 (4.95 - 10.21)	0.40 (0.22 - 0.58)**	2.45 (1.47 - 3.45)	2.08 (1.69 - 2.47)	20%	0.11 (0.04 - 0.19)
October (Baltimore only)	56	18.77 (6.34 - 31.20)	0.79 (0.57 - 1.02)	5.33 (1.86 - 8.80)	2.47 (1.80 - 3.15)	20%	0.65 (0.12 - 1.19)
Metropolitan area							
Baltimore, MD	102	13.73 (6.81 - 20.64)	0.62 (0.47 - 0.78)	4.04 (2.08 - 5.99)	2.30 (1.89 - 2.70)	20%	0.41 (0.12 - 0.70)
Denver, CO	40	5.30 (4.05 - 6.54)	0.31 (0.27 - 0.36)	1.46 (0.96 - 1.96)	2.42 (1.94 - 2.91)	10%	0.053 (0 - 0.11)
Los Angeles, CA	42	10.59 (6.81 - 14.36)	0.40 (0.34 - 0.46)	2.39 (1.45 - 3.33)	2.51 (1.99 - 3.02)	17%	0.13 (0.027 - 0.23)
Processor							
Producer A	6	12.68 (4.02 - 21.34)	1.71 (0.41 - 3.01)	1.35 (-0.06 - 2.77)	1.80 (1.08 - 2.52)	17%	0.30 (0.48 - 1.08)
Producer B	34	31.56 (11.64 - 51.47)	0.77 (0.39 - 1.16)	9.17 (3.57 - 14.77)	4.13 (3.17 - 5.10)	44%	1.09 (0.24 - 1.95)
Producer C	14	3.71 (0.98 - 6.44)	0.36 (0.28 - 0.44)	0.90 (0.32 - 1.48)	1.59 (0.40 - 2.79)	21%	0.13 (0 - 0.27)
Producer D	14	6.70 (5.61 - 7.80)	0.61 (0.47 - 0.74)	1.47 (0.98 - 1.95)	2.88 (2.23 - 3.53)	0%	NA
Producer E	30	0.98 (0.82 - 1.15)	0.33 (0.29 - 0.37)	0.28 (0.24 - 0.32)	0.76 (0.65 - 0.86)	0%	NA
Producer F	27	5.24 (4.13 - 6.35)	0.28 (0.24 - 0.32)	1.34 (0.79 - 1.90)	1.95 (1.56 - 2.33)	7%	0.068 (0 - 0.16)
Producer G	20	4.54 (3.24 - 5.84)	0.39 (0.34 - 0.45)	1.72 (1.22 - 2.22)	1.80 (1.36 - 2.25)	10%	0.053 (0 - 0.13)
Producer H	8	32.87 (25.42 - 40.32)	0.47 (0.36 - 0.57)	7.86 (6.05 - 9.67)	4.12 (3.05 - 5.19)	75%	0.63 (0.20 - 1.06)
Producer I	6	8.14 (5.37 - 10.91)	0.43 (0.32 - 0.53)	3.59 (2.41 - 4.78)	1.86 (1.43 - 2.28)	0%	NA
Producer J	6	6.29 (4.87 - 7.70)	0.33 (0.20 - 0.46)	1.60 (0.58 - 2.62)	2.25 (1.93 - 2.57)	0%	NA
Other Producers	19	7.91 (5.72 - 10.09)	0.45 (0.36 - 0.54)	2.34 (1.48 - 3.20)	2.82 (2.38 - 3.26)	10%	0.04 (0 - 0.10)

Detection limits were reported in dry weight as 1, 1 - 2, 1, 1, 1 $\mu\text{g As kg}^{-1}$ for total arsenic, iAs, MA, DMA and nitarsonic, respectively.

* $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$; NA: Not applicable, all samples below the detection limit; Sample-specific concentration estimates were derived by taking the average of the replicate measurements from the same package (2 - 3 replicates for total arsenic and each arsenic species). For total arsenic, iAs, MA and DMA measurements below the detection limit, we imputed the value of the detection limit divided by the square root of two. For nitarsonic, samples below the detection limit were given the value of 0.

Table 2. Estimates of lifetime average daily dose of arsenic species in adults and average daily dose in children (in $\mu\text{g kgBW}^{-1} \text{ day}^{-1}$) resulting from consumption of turkey, based on sample characteristics

Adult

Sample characteristic	iAs	MA	DMA	Nitarsone
Antibiotic-free or USDA-Certified Organic	0.00019	0.00075	0.00106	0.00003
Conventional, all	0.00028	0.00196	0.00122	0.00019
Conventional with prohibitory policy	0.00016	0.00014	0.00037	NA
Conventional, no known arsenical policy	0.00031	0.00260	0.00146	0.00026
No nitarsone detection	0.00020	0.00074	0.00095	NA
Positive nitarsone detection	0.00045	0.00537	0.00222	0.00078

Child (aged 4 - 30 months)

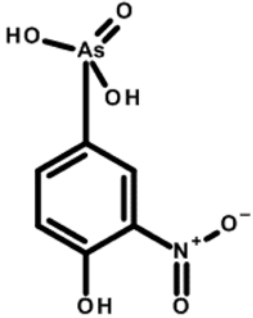
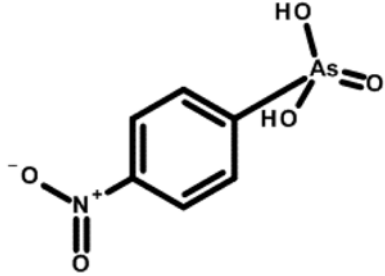
Sample characteristic	iAs	MA	DMA	Nitarsone
Antibiotic-free or USDA-Certified Organic	0.00073	0.00280	0.00390	0.00010
Conventional, all	0.00105	0.00720	0.00450	0.00073
Conventional with prohibitory policy	0.00061	0.00051	0.00150	NA
Conventional, no known arsenical policy	0.00119	0.00950	0.00560	0.00093
No nitarsone detection	0.00077	0.00270	0.00350	NA
Positive nitarsone detection	0.00171	0.02039	0.00820	0.00300

NA: Not applicable, all samples below the detection limit. Arsenic species mean concentration estimates from Table 1 were used to calculate daily doses. Turkey meat intake rates of 0.49 and 1.86 $\text{g kgBW}^{-1} \text{ day}^{-1}$, calculated from NHANES dietary intake data, were used for adults and children aged 4 – 30 months, respectively, in estimation of daily doses. Body weights of 80 and 11 kg (estimated from NHANES data) were used for adults and children aged 4 – 30 months, respectively.

FIGURE LEGENDS

Figure 1. Comparison of two arsenical poultry drugs, roxarsone and nitarsonsone. Dosage rate and indication information are from the Food and Drug Administration's Animal Drugs @ FDA website (<http://www.accessdata.fda.gov/scripts/animaldrugsatfda/>). Roxarsone: <http://www.accessdata.fda.gov/scripts/animaldrugsatfda/details.cfm?dn=007-891>; Nitarsonsone: <http://www.accessdata.fda.gov/scripts/animaldrugsatfda/details.cfm?dn=007-616>.

Figure 1. Comparison of two arsenical poultry drugs, roxarsone and nitarosone

Structure	<p style="text-align: center;">Roxarsone</p> 	<p style="text-align: center;">Nitarosone</p> 
Species	Broiler chicken	Turkey
Dosage rate	22.7 - 45.5 g ton ⁻¹	170.0 - 187.5 g ton ⁻¹
Purpose (indication) for use	Improved feed conversion, weight gain and pigmentation; prevention of coccidiosis	Prevention of blackhead disease
US FDA approval status	Withdrawn (February 2014)	Withdrawn (December 2015)